

1 **Contrasting water use characteristics of riparian trees under different**
2 **water tables along a losing river**

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23 **Abstract**

24 Rivers losing flow into surrounding aquifers ('losing' rivers) are common under changing
25 climates and groundwater overexploitation. The riparian plant-water relations under various
26 water table dynamics along a losing river remain unclear. In this study, the water isotopes
27 ($\delta^2\text{H}$ and $\delta^{18}\text{O}$), leaf $\delta^{13}\text{C}$, and MixSIAR model were used combinedly for determining the
28 root water uptake patterns and leaf water use efficiency (WUE) of *Salix babylonica* (L.) at
29 three sites (A, B, and C) with different water table depths (WTDs) in the riparian zone of Jian
30 and Chaobai River in Beijing, China. The correlations of water source contributions with
31 WTD and WUE were quantified. The riparian *S. babylonica* primarily took up upper (0–80
32 cm) soil water (71.5%) with the lowest leaf $\delta^{13}\text{C}$ ($-28.8 \pm 1.1 \text{‰}$) at site A under deep WTD
33 ($20.5 \pm 0.5 \text{ m}$). In contrast, deep water below 80 cm depth including groundwater contributed
34 55.1% to *S. babylonica* at site B with fluctuated shallow WTD ($1.9 \pm 0.4 \text{ m}$), where leaf $\delta^{13}\text{C}$
35 was highest ($-27.9 \pm 1.0 \text{‰}$). The *S. babylonica* mainly used soil water in 30–170 cm layer
36 (56.9%) with mean leaf $\delta^{13}\text{C}$ of $-28.2 \text{‰} \pm 0.7 \text{‰}$ at site C with stable shallow WTD ($1.5 \pm$
37 0.1 m). It was found that both the contributions of upper soil water in 0–80 cm and deep
38 water below 80 cm had significantly quadratic correlations with WTD under shallow water
39 table conditions ($p < 0.05$). Leaf $\delta^{13}\text{C}$ was negatively correlated with contributions of upper
40 soil water above 80 cm depth, but it was positively related to the contributions of deep water
41 below 80 cm in linear functions ($p < 0.001$). The results indicated that 2.1 m was the
42 optimum WTD for riparian trees, because they maximized the use of deep water sources to
43 get the highest WUE. This study provides insights into managing groundwater, surface water
44 resources and riparian afforestation in losing rivers.

45 **Key words:** Root water uptake pattern; Water use efficiency; Water table depth; Stable
46 isotopes; Riparian *S. babylonica*; Losing river

47

48 **1. Introduction**

49 Many rivers are drying up under warming climates and groundwater overexploitation
50 worldwide, which causes serious decline of water table and degradation of riparian vegetation
51 (Rood et al., 2003; Garcia et al., 2017). Ecological water (e.g., excess surface flow or
52 reclaimed water) has been widely used to recharge dry rivers in recent years (Long et al.,
53 2020; He et al., 2021). Widespread loss of riverflow into underlying aquifers occurs where
54 groundwater levels lie below nearby rivers ('losing' rivers) (Winter et al., 1998; Jasechko et
55 al., 2021). The recharge process increases the water availability for riparian vegetation and
56 various species have been planted to restore the riparian ecosystem. However, increased plant
57 transpiration was linked to reduced riverflow and groundwater level (Vanderklein et al., 2014;
58 Missik et al., 2019; Mkunyana et al., 2019), although some studies claimed that transpiration
59 was not closely coupled to riverflow (McDonald et al., 2015). Therefore, investigating the
60 water use characteristics of riparian species is critical to recover the losing rivers especially in
61 water-scarce regions. It will provide critical insights into managing groundwater and surface
62 water resources as well as riparian afforestation under climate change.

63 The $\delta^2\text{H}$ and $\delta^{18}\text{O}$ have been widely applied to identify the root water uptake patterns of
64 riparian plants (Dawson and Ehleringer, 1991; Wang et al., 2019a; Oerter and Bowen, 2019;
65 Flanagan et al., 2019). Dawson and Ehleringer (1991) utilized deuterium tracer technique to
66 identify that mature trees nearby a perennial stream rarely absorbed the stream water but used

67 water from deeper strata in the Wasatch Mountains of semiarid western USA. However,
68 Oerter and Bowen (2019) revisited the same experimental site in Dawson and Ehleringer
69 (1991) and demonstrated that soil water was the primary source for these streamside trees via
70 the *in situ* soil water vapor isotope measurements. Applying the stable isotope ($\delta^2\text{H}$ and $\delta^{18}\text{O}$)
71 tracer technique, several studies concluded that the distances away from the riverbank as well
72 as seasonal variations of climate could significantly affect water uptake patterns of riparian
73 trees. For example, riparian trees at distances greater than 5–15 m from the riverbank
74 substantially depended on soil water and groundwater for plant transpiration without water
75 absorption from the nearby river water throughout the growing season (Mensforth et al., 1994;
76 Qian et al., 2017; Wang et al., 2019a). Shallow soil water recharged by recent precipitation
77 was generally used as the primary water source for riparian plants during wet season, whereas
78 deep soil water and groundwater were principal in dry season (Dawson and Pate, 1996; Wang
79 et al., 2019a). In particular, groundwater is perceived as a reliable and continuous water
80 source for deep-rooted riparian trees indicated by consistent water isotopes of stems and
81 groundwater (Barbeta and Penuelas, 2017; Li et al., 2019).

82 The water table depth (WTD) has great effects on riparian plant water use (Nippert et al.,
83 2010; Chen et al., 2016; Jasechko et al., 2021). Chen et al. (2016) found that desert riparian
84 species under the WTD of 1.8 m used a mixture of soil water below 75 cm depth,
85 groundwater, and river water. Once the WTD increased to 3.8 m and 7.2 m, species mainly
86 relied on deeper soil water below 150 cm and even groundwater. Majority of studies
87 indicated that the increasing WTD induced water uptake pattern adjustments towards deeper
88 water sources (Barbeta et al., 2015; Chen et al., 2016; Li et al., 2019). Nevertheless, some

89 studies argued that riparian trees switched their water sources from groundwater to
90 vadose-zone soil water in response to the water table decline (Nippert et al., 2010; Busch et
91 al., 1992). Riparian species might change root distributions and develop plastic water uptake
92 patterns in order to acclimate to the variable WTDs (Horton et al., 2001). Therefore, it is
93 necessary to investigate the water uptake patterns for riparian trees under various water table
94 conditions. Furthermore, little is known about the quantitative relationship between the WTD
95 and root water uptake patterns of the riparian plants.

96 The shift of water sources for riparian plants is mainly driven by leaf transpiration. Leaf
97 water use efficiency (WUE), defining as the ratio of photosynthetic rate and transpiration rate,
98 is a crucial indicator to reflect plant water use characteristics (Farquhar et al., 1989; Wang et
99 al., 2021; Zhao et al., 2021). The plant leaf $\delta^{13}\text{C}$ is positively correlated with intrinsic WUE
100 especially for C3 plants and it can be used as an indicator of WUE (Wang et al., 2019b). A
101 growing numbers of evidence showed that plants with higher dependence on deeper water
102 sources were generally linked to higher WUE and survive drought more successfully
103 (Picon-Cochard et al., 2001; Liu et al., 2017; Ding et al., 2020). The WUE of arid and
104 semi-arid deep-rooted plant species was observed to increase with deeper WTD (Si et al.,
105 2015), whereas Antunes et al. (2018) found that there was no response of WUE to increasing
106 WTD for the semi-arid dimorphic-rooted conifer trees. It can be found that most of above
107 studies are located in arid and semi-arid areas. It is still uncertain how WUE of riparian plants
108 coordinates with root water uptake patterns under different WTD conditions in losing rivers.

109 The aim of this study was to identify and understand water strategies of deep-rooted
110 riparian trees along gradients of WTD, using stable isotopes ($\delta^2\text{H}$, $\delta^{18}\text{O}$ and $\delta^{13}\text{C}$) together

111 with the MixSIAR model. Focusing on rivers in Beijing, China, the objectives of this study
112 were (1) to quantify the seasonal water uptake patterns for riparian trees under different
113 WTDs and distances away from the riverbank, (2) to compare the leaf WUE of riparian trees
114 under different WTDs and distances away from the riverbank, (3) to determine the
115 quantitative relationships of water uptake patterns with WTD and leaf WUE. These results
116 will provide insights on riparian ecosystem and water resources management of losing rivers.

117 **2. Materials and methods**

118 *2.1 Study area and field measurements*

119 The experiments of *Salix babylonica* (L.) were conducted from May to November in 2019
120 in the riverside of the Jian and Chaobai River, Beijing, China (40°07'30"N, 116°40'37"E) (Fig.
121 1). A sub-humid monsoon climate with mean temperature of 11.5 °C and mean relative
122 humidity (RHU) of 60% is dominant in this study area. The multi-year (1961–2021) average
123 precipitation is 532 mm, with 84.5% of which occurring at the rainy season during June 15 and
124 September 15 (Fig. 2). In comparison, the total precipitation in 2019 was 445.6 mm, with
125 73.9% of which concentrated on the rainy season (Fig. 2). Thus, 2019 was identified as a dry
126 year based on the multi-year (1961–2021) rainfall frequency curve. The Jian and Chaobai
127 River were dried up between 1999 and 2007 and have been recharged with ecological water
128 since 2007. *S. babylonica* with deep roots was one of the most widespread planted riparian
129 species in the study area. The maximum rooting depth of *S. babylonica* can reach about 4 m
130 and vary with different water tables (Ferro et al., 2003). Three representative sites (A, B, and C)
131 with different WTDs were selected along the Jian and Chaobai River (Fig. 1). At each site, four
132 plots planted with *S. babylonica* at distance of 5 m, 10 m, 20 m and 45 m away from the

133 riverbank (D05, D10, D20 and D45, respectively) were constructed for field measurements
134 and sample collections (Fig. 1).

135 <Figure 1>

136 A tilting rain gauge of 0.2 mm per tip (SL3-1, Shanghai meteorological instrument,
137 Shanghai, China) was used to record daily precipitation in the meteorological observation
138 station (Fig. 1). The river water level nearby each site was monitored using a doppler open
139 channel flowmeter (HOH-L-01, King Water Co Ltd., Beijing, China). The groundwater level
140 was measured once a month using a pressure-type groundwater stage gauge (HOH-S-Y, King
141 Water Co Ltd., Beijing, China) placed in the monitoring wells. The river water level was
142 observed to fluctuate at 28.4 – 29.4, 27.9 – 28.9, and 26.5 – 27.3 m nearby the sites A, B, and C,
143 respectively (Fig. 2a). The WTD at site A ($20.5 \text{ m} \pm 0.5 \text{ m}$) was significantly ($p < 0.01$) deeper
144 than the other two sites B ($1.9 \text{ m} \pm 0.4 \text{ m}$) and C ($1.5 \text{ m} \pm 0.1 \text{ m}$). The WTD fluctuation was
145 more significant at site B compared with the stable WTD at site C (Fig. 2b).

146 <Figure 2>

147 2.2 Sample collection

148 Six field campaigns were conducted to collect water samples of river, groundwater,
149 precipitation, soil, and stem for $\delta^2\text{H}$ and $\delta^{18}\text{O}$ measurements as well as plant leaf samples for
150 $\delta^{13}\text{C}$ analysis on May 5, June 14, July 26, August 15, September 26, and November 5. River
151 water nearby each site was sampled by means of a hydrophore, while a sucking pump was
152 used to collect groundwater in monitoring wells. Precipitation was sampled using a device
153 consisted of a polyethylene bottle, a funnel and plastic ball during the observation period
154 (Sun et al., 2019).

155 One riparian *S. babylonica* (with diameter of 23 ± 3 cm at breast height) in each plot was
156 selected to measure the water stable isotopes in stem and leaf $\delta^{13}\text{C}$ (Fig. 1). Several suberized
157 and mature stems were sampled from each *S. babylonica* and removed the phloem and bark to
158 prevent isotopic contamination. Stems were put into 12-ml glass vials and kept refrigerated
159 before stem water extraction and isotope analysis. Mature foliage (approximately 50 leaves)
160 from branches at similar heights (4–5 m) of *S. babylonica* tree was collected for plant leaf $\delta^{13}\text{C}$
161 analysis. The collected leaves were oven-dried at $65\text{ }^\circ\text{C}$ for at least 64 h until they reached a
162 constant mass on the day of sampling. Subsequently, leaf samples were ground into fine
163 powder and sieved using a 0.15 mm mesh screen.

164 Soil cores within 1 m of each *S. babylonica* tree in the four plots were obtained by a power
165 auger (CHPD78, Christie Engineering Company, Sydney, Australia). The soil samples were
166 collected at depths of 0–5, 5–10, 10–15, 15–20, 20–30, 30–40, 40–60, 60–80, 90–110, 150–
167 170, 190–210, 250–270, and 280–300 cm. They were removed roots, put into 12-ml glass vials
168 and kept frozen in a refrigerator before soil water extraction and isotope analysis. Soil was also
169 sampled to measure the gravimetric soil water content (SWC) using the oven-dry method at the
170 same depths listed above.

171 2.3 Isotopic analyses

172 Water in stem and soil samples was extracted using the automatic cryogenic vacuum
173 distillation system (LI-2100, LICA, Beijing, China). The whole extraction progress lasted for 3
174 h and the extraction efficiency was all up to 99% to ensure no isotopic fractionation. The
175 isotopic ratio infrared spectroscopy (IRIS) system (DLT-100, Los Gatos Research, Mountain
176 View, USA) was applied to measure the $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in precipitation, river water, soil water,

177 and groundwater. To avoid the organic contaminants during IRIS measurements, the isotopic
178 composition of stem water was analyzed using the Isotope Ratio Mass Spectrometry (IRMS)
179 system (MAT253, Thermo Fisher Scientific, Bremen, Germany). The measuring precisions of
180 both IRIS and IRMS system were $\pm 1\%$ for $\delta^2\text{H}$ and $\pm 0.1\%$ for $\delta^{18}\text{O}$ in different water samples,
181 respectively. The measured $\delta^2\text{H}$ and $\delta^{18}\text{O}$ were then calibrated with the Vienna Standard Mean
182 Ocean Water (VSMOW) (Sun et al., 2019; Zhao and Wang, 2021). There was no significant
183 difference in $\delta^2\text{H}$ and $\delta^{18}\text{O}$ measurements for soil water, groundwater, precipitation and river
184 water analyzed by the IRMS and IRIS systems (Li et al., 2021). The $\delta^{13}\text{C}$ in plant leaf was
185 measured using the IRMS system (MAT253, Thermo Fisher Scientific, Bremen, Germany)
186 with a precision of $\pm 0.15\%$. The measured $\delta^{13}\text{C}$ in plant leaf was calibrated with the Vienna
187 Pee Dee Belemnite (V-PDB).

188 2.4 MixSIAR model

189 The MixSIAR model v3.1 (Stock and Semmens, 2016) coupled with $\delta^2\text{H}$ and $\delta^{18}\text{O}$ isotopes
190 was applied to identify the potential water source contributions to riparian *S. babylonica*.
191 Because the WTD was much deeper than the maximum rooting depth of *S. babylonica* trees at
192 site A, groundwater could not be taken up by trees. In light of the water isotopic composition
193 and SWC distributions in the profile, soil water in the four layers of 0–30, 30–80, 80–170, and
194 170–300 cm were classified as the potential water sources of *S. babylonica* trees at site A.
195 Since water isotopic values of groundwater were similar to those of soil water in the 170–300
196 cm layer, they had been merged into one potential water source (soil water in 170–300 cm
197 layer) besides the soil water sources in the 0–30 cm, 30–80 cm, and 80–170 cm layers for *S.*
198 *babylonica* trees under shallow WTD sites B ($1.9\text{ m} \pm 0.4\text{ m}$) and C ($1.5\text{ m} \pm 0.1\text{ m}$). These

199 isotopic compositions of potential water sources were referred as source data in the MixSIAR
200 model, while the mixture data was set as the $\delta^2\text{H}$ corrected by potential water source line (PWL)
201 correction method (Li et al., 2021) and raw $\delta^{18}\text{O}$ in stem water. More details about the
202 MixSIAR model parameter settings were given in Li et al. (2021).

203 *2.5 Statistical analysis*

204 The seasonal differences in the isotopes of different water bodies, SWC, proportional
205 contributions of water sources, and plant leaf $\delta^{13}\text{C}$ values among the three sites as well as those
206 among the four plots at different distances away from the riverbank were evaluated by the
207 One-way analysis of variance (ANOVA) incorporating with post-hoc Tukey's,
208 Kolmogorov-Smirnov and Levene's tests ($p < 0.05$) (Sun et al., 2018). Regression analyses
209 were performed to quantify the relationships of water source contributions with WTD and
210 plant leaf $\delta^{13}\text{C}$ values. All statistical analyses were conducted with the SPSS 24.0 software
211 (SPSS Inc., USA).

212 **3. Results**

213 *3.1 Variations of soil water contents and isotopes in different waters*

214 The SWC in different soil layers varied greatly with sampling campaigns and distances
215 away from the riverbank among the sites A, B and C (Fig. 3). The SWC values in upper 0–80
216 cm layer were significantly higher at site A than other two sites throughout the observation
217 period ($p < 0.001$). However, the SWC in 170–300 cm layer at site A was much lower with
218 significantly seasonal variations compared with other two sites ($p < 0.001$). A significant
219 decrease of SWC was observed from May (0.32 g g^{-1}) to July (0.09 g g^{-1}) in 170–300 cm
220 layer at site A. The SWC in 80–170 cm layer was lower but changed greatly with seasons

221 (standard deviation of 0.8 g g^{-1}) at site B than other two sites. Furthermore, no significant
222 difference in SWC in 0–80 cm layer was found among four distances at site A ($p > 0.05$), but
223 decreasing SWC was found in 170–300 cm layer from D05 to D45 (Fig. S1). The SWC in
224 80–170 cm layer decreased from D05 to D45 at site B. There was no significant difference in
225 SWC distributions among the four distances at site C ($p = 0.98$) (Fig. S1).

226 <Figure 3>

227 <Figure S1>

228 The isotopes in groundwater at site A (mean of -71.1‰ for $\delta^2\text{H}$ and -10.2‰ for $\delta^{18}\text{O}$) were
229 significantly more depleted than those of groundwater at sites B (mean of -58.9‰ for $\delta^2\text{H}$ and
230 -7.6‰ for $\delta^{18}\text{O}$) and C (mean of -59.4‰ for $\delta^2\text{H}$ and -7.7‰ for $\delta^{18}\text{O}$) ($p < 0.01$) (Fig. 4 and
231 Table S1). There was no significant difference in the isotopic compositions of river water
232 among the three sites ($p > 0.05$). It was obvious that the $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in river water were much
233 more enriched than those in groundwater at all the three sites ($p < 0.05$) (Fig. 4 and Table S1).

234 The isotopic compositions of soil water had no significant difference in the 0–80 cm layer
235 among the three sites ($p = 0.96$), but they were more enriched below 80 cm at sites B and C
236 compared with those at site A ($p < 0.05$). The slopes of soil water line (SWL) at sites B and C
237 were much lower than those at site A, indicating stronger soil water evaporation at sites B and
238 C. No significant difference in isotopic values of soil water was observed among four plots
239 (D05, D10, D20, D45) at site A ($p = 0.83$). The $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in soil water in the 80–170 cm
240 layer were gradually depleted from D05 to D45 during June and August at site B, while they
241 increased from D05 to D20 but then decreased in plot D45 at site C (Fig. 4 and Table S1).

242 The slope of SWL gradually increased with the increase of distance away from the riverbank at

243 the three sites, suggesting a decrease of soil water evaporation from D05 to D45. The stem
244 water isotopes at sites A and C were slightly depleted than those at site B (Fig. 4 and Table S1).
245 $\delta^2\text{H}$ and $\delta^{18}\text{O}$ in stem water gradually enriched from D05 to D20 at sites A and C, whereas they
246 were depleted from D05 to D45 at site B.

247 <Figure 4>

248 3.2 Seasonal and spatial variations in root water uptake patterns

249 The water uptake patterns of riparian *S. babylonica* were strongly different among the
250 three sites (Figs. 5 and 6). *S. babylonica* primarily took up upper soil water in 0–80 cm layer
251 with mean contributions of 71.5% throughout the observation period at site A with deep
252 WTD (Figs. 5a and 6a). The soil water below 80 cm depth primarily contributed to *S.*
253 *babylonica* with proportions more than 55.1% at site B with fluctuated shallow WTD (Figs.
254 5b and 6b). In comparison, the main water source of *S. babylonica* at site C under shallow
255 WTD without significant fluctuation was soil water in 30–170 cm layer with mean
256 contributions of 56.9% throughout the observation period (Figs. 5c and 6c). *S. babylonica* at
257 site A absorbed particularly more soil water in 0–30 cm layer (44.2%) than sites B (20.7%)
258 and C (22.8%) ($p < 0.01$). Moreover, the proportional contributions of soil water in 170–300
259 cm layer or groundwater to *S. babylonica* were significantly different among sites A (13.6%),
260 B (28.3%), and C (18.3%) ($p < 0.01$) (Fig. 6).

261 The contributions of potential water sources to *S. babylonica* exhibited significantly
262 seasonal differences among the three sites ($p < 0.05$). A decrease of the soil water
263 contribution below 80 cm depth associated with an increase in 0–80 cm layer was shown
264 with the growth of *S. babylonica* trees at site A (Fig. 5a). The soil water contribution in

265 80–170 cm layer to *S. babylonica* decreased significantly (21.6%) during June to August with
266 a water table decline (0.8 m) at site B (Fig. 5b) ($p < 0.05$). Meanwhile, *S. babylonica* water
267 uptake from surficial soil water in 0–30 cm layer increased by 20.0% during rainy season ($p <$
268 0.05). *S. babylonica* trees at site C kept reducing the water uptake proportions in 30–80 cm
269 layer from May (46.7%) to November (17.7%), while groundwater contributions increased by
270 10.7% (Fig. 5c) ($p < 0.05$). Although significant seasonal differences in riparian tree water
271 uptake depths were observed ($p < 0.05$), there was no general and significant correlation
272 between environmental factors (e.g., air temperature, RHU, and precipitation) and upper soil
273 water (0–80 cm) contribution to riparian trees at the three sites ($p > 0.05$) (Fig. S2).

274 The water use characteristics of *S. babylonica* varied distinctly along the distance away
275 from the riverbank at the three sites. The superficial (0–30 cm) soil water contributions
276 increased significantly by 23.0% from D05 to D20 ($p < 0.05$), whereas deep (170–300 cm)
277 soil water contributions to *S. babylonica* reduced by 6.4% from D05 to D45 at site A (Fig. 6a).
278 The *S. babylonica* took up more groundwater but less soil water in 80–170 cm layer from
279 D05 to D45 at site B (Fig. 6b). In particular, the groundwater contribution to *S. babylonica* at
280 D20 and D45 significantly increased by an average of 14.4% when the water table declined
281 from 1.7 m to 2.5 m at site B ($p < 0.05$). On the contrary, *S. babylonica* reduced water
282 absorption (47.3%) from deep water below 80 cm but accessed more upper soil water (28.6%)
283 significantly ($p < 0.05$) in response to the 0.8 m decline of WTD at D05 and D10 at site B (p
284 < 0.05). At site C, the contributions of deep water below 80 cm depth reduced remarkably
285 (20.2%) from D05 to D45, whereas the upper soil water increased the contributions (14.1%) to
286 *S. babylonica* from D05 to D20 (Fig. 6c) ($p < 0.05$).

287 <Figure 5>

288 <Figure 6>

289 3.3 Changes in leaf $\delta^{13}\text{C}$ of riparian trees

290 The leaf $\delta^{13}\text{C}$ values at the three sites varied differently with increasing distance from the
291 riverbank during the observation period (Fig. 7). Riparian trees at site A had significantly
292 lower leaf $\delta^{13}\text{C}$ values (mean of -28.8‰) than those at the other two sites B (mean of
293 -27.9‰) and C (mean of -28.1‰) ($p < 0.05$). There was a significant decrease of leaf $\delta^{13}\text{C}$
294 value (2.3‰) from D05 to D20 at site A ($p < 0.05$). However, the $\delta^{13}\text{C}$ values generally
295 increased with increasing distance from the riverbank (ranging from -28.8‰ at D05 to -27.0‰
296 at D45) at site B. It was found that the leaves of *S. babylonica* at D20 and D45 had
297 significantly higher $\delta^{13}\text{C}$ values (mean of -27.1‰) than those at other two distances (mean of
298 -28.7‰) at site B ($p < 0.05$). The leaf $\delta^{13}\text{C}$ of the *S. babylonica* nearest the riverbank (mean
299 of -27.6‰ at D05) at site C was higher than that at other distances (mean of -28.4‰) (Fig.
300 7). Significant seasonal differences in the leaf $\delta^{13}\text{C}$ values were shown among the three sites
301 ($p < 0.05$) (Table S2). The $\delta^{13}\text{C}$ values in the tree leaves were the lowest in May (mean of
302 -28.7‰) and August (mean of -28.7‰). The leaf $\delta^{13}\text{C}$ in plot D05 at site C showed an
303 increasing trend with the growth of *S. babylonica* compared with that at sites A and B (Table
304 S2). Nevertheless, no significant regression relationship between leaf $\delta^{13}\text{C}$ value of *S.*
305 *babylonica* and environment factors (e.g., air temperature, RHU, and precipitation) was
306 observed at the tree sites ($p > 0.05$) (Fig. S3).

307 <Figure 7>

308 **4. Discussion**

309 *4.1 Impacts of water table depth on riparian tree water uptake pattern*

310 This study indicated that the depth and fluctuation of water table diversified the water
311 uptake patterns of riparian trees near the losing river (Naumburg et al., 2005; Simunek and
312 Hopmans, 2009; Fan et al., 2017; Antunes et al., 2018). It agreed well with previous studies
313 that riparian trees primarily took up soil water in upper 0–80 cm layer under deep WTD with
314 weaker hydraulic connections between the river and underlying groundwater system (Figs. 5a
315 and 6a) (Wu et al., 2021). An evident shift of the main water uptake depth of riparian trees
316 from soil water in 30–80 cm layer towards deeper soil layers or groundwater was found when
317 the WTD increased from 1.5 ± 0.1 m to 1.9 ± 0.4 m under shallow WTD condition with
318 closer river water-groundwater interactions in this study (Figs. 5b, 5c, 6b and 6c). It
319 suggested that the deep-root growth and their water uptake capabilities were limited under
320 shallow WTD without significant fluctuations (1.5 ± 0.1 m) (Figs. 5c and 6c) or oxygen stress
321 conditions (Naumburg et al., 2005; Fan et al., 2017). Nevertheless, riparian trees could
322 gradually acclimate to persistently shallow WTD environment with plant growth indicated by
323 an increase of groundwater utilization (10.7%) for riparian trees from May to November
324 (Figs. 5c and 6c) (Naumburg et al., 2005). In addition, these riparian trees closer to the stream
325 generally had better ability to adapt to oxygen stress, which could be derived from the
326 decreasing deep-water contributions along the distance away from riverbank (Fig. 6c).
327 Fluctuated shallow WTD could motivate riparian trees to develop a dimorphic-root system
328 and switch their water main uptake depths between the shallow soil layers and deep layers
329 (Figs. 5b and 6b) (Zencich et al., 2002). But the responses to water table decline of riparian

330 trees at different distances at fluctuated WTD area differed significantly with reducing water
331 absorption from deep water below 80 cm at distance within 10 m and increasing groundwater
332 utilization at distance of 20 m and 45 m away from the riverbank (Figs. 5b and 6b). The
333 stimulation of deep-roots helped riparian trees susceptible to drought stress at distance greater
334 than 20 m from the riverbank survive from the climate drought (Ehleringer and Dawson,
335 1992; Antunes et al., 2018; Li et al., 2019) or river water level lowering during rainy season
336 as shown in this study (Figs. 3c and 5b).

337 The quantitative relationship between the shallow WTD and contributions of different
338 water sources to riparian trees was not discovered in previous studies. In our study, a
339 significantly quadratic relationship was found between the WTD and proportional
340 contributions of upper soil water in 0–80 cm layer ($R^2 = 0.43$, $p < 0.05$) at shallow WTD
341 areas (Sites B and C) (Fig. 8a). Meanwhile, the WTD had a significant quadratic correlation
342 with deep-water contributions below 80 cm depth ($R^2 = 0.43$, $p < 0.05$) (Fig. 8b). Particularly,
343 riparian *S. babylonica* at constant shallow WTD sites (WTD < 1.7 m) increased deep water
344 absorption with increasing WTD, whereas those under fluctuated shallow WTD (1.7–2.5 m)
345 showed an increase and then decrease of deep-water contribution as water table declined (Fig.
346 8b). The upper soil water contribution was minimum at the WTD of 2.1 m, when the
347 deep-water contribution reached a maximum value. It could be explained that declining water
348 tables (WTD increased from 1.5 m to 2.1 m) aerated deep soil layers, benefited for new
349 deep-root exploitation and further stimulated riparian trees to utilize deep water when
350 groundwater severely restricted the active root zone of riparian trees under oxygen stress
351 (Naumburg et al., 2005). However, riparian trees can tolerate declining water table only up to

352 a species-specific threshold value, beyond which drought stress occurred (Sperry et al., 1998).
353 In this study, when the WTD exceeded 2.1 m, deeper water tables could reduce the
354 availability of a permanent water source for riparian trees and therefore lead to water stress
355 (Naumburg et al., 2005; Wang et al., 2019a). Therefore, water stress and oxygen stress
356 resulting from fluctuated WTD had a great effect on riparian tree water uptake patterns.
357 Specifically, higher water tables without evident fluctuations generally killed flooded roots
358 because most species cannot tolerate the associated low oxygen levels (Naumburg et al., 2005;
359 Li et al., 2019). The appropriately decreasing water tables benefited plant species from
360 avoiding oxygen stress, but the excessive water table decline probably resulted in plant water
361 stress (Naumburg et al., 2005; Li et al., 2019). Our findings suggested that 2.1 m was the
362 optimum WTD to utilize the deep water sources for the riparian trees development in the face
363 of drought and oxygen scarce environment.

364 <Figure 8>

365 *4.2 Relationship between riparian tree water uptake depth and water use efficiency*

366 It was evident that there were tradeoff and coordination between plant water uptake depth
367 and WUE near rivers, because significant difference of WUE was observed in riparian trees
368 among the four plots at different distances from the riverbank at three sites with various root
369 water uptake patterns (Figs. 5, 6 and 7). Plants with shallow root water uptake depths (such as
370 *S. babylonica* at deep WTD at site A) mainly relied on upper soil water which was subjected
371 to fast evaporation loss (Figs. 5a and 6a). They generally had rapid water resource capture
372 and large transpiration rates (Ding et al., 2020; Nie et al., 2019; Antunes et al., 2018), leading
373 to lower leaf $\delta^{13}\text{C}$ and WUE values (Figs. 5a and 7). Moreover, the increase in shallow soil

374 water contributions was linked to the obviously decreasing leaf $\delta^{13}\text{C}$ and WUE values as the
375 distance from the riverbank increased (from 5 m to 20 m) at deep WTD site (Figs. 5a and 7).
376 When the riparian trees primarily depended on deep water sources in fluctuated shallow
377 WTD area (such as *S. babylonica* at site B), leaf $\delta^{13}\text{C}$ and WUE increased significantly
378 particularly at the lowering of the water table (Figs. 5b and 7). The increasing deep water
379 contributions accompanied with growing leaf $\delta^{13}\text{C}$ and WUE were further found along the
380 distance away from riverbank at high water table site with evident fluctuations. This could be
381 likely due to that stomatal conductance and transpiration rate reduced with the decreasing
382 deep soil moisture to avoid hydraulic failure and drought stress (Klein et al., 2013;
383 Martinez-Vilalta and Garcia-Forner, 2017; Ding et al., 2020). Pezeshki et al. (1998) and Li et
384 al. (2006) also indicated that the anoxia could reduce the plant stomatal conductance and
385 transpiration rate. Our findings confirmed that deep-root water uptake and leaf transpiration
386 rate of riparian trees in anaerobiosis environment were restricted under the constant shallow
387 WTD condition (Figs. 5c and 7).

388 The relationship between the riparian tree water uptake depth and WUE was not
389 quantified in previous studies. As shown in Fig. 9a, this study clarified that the leaf $\delta^{13}\text{C}$ was
390 significantly negatively correlated with the contributions of upper soil water in 0–80 cm layer
391 ($R^2 = 0.47, p < 0.001$). Meanwhile, significantly positive relationships were found between
392 leaf $\delta^{13}\text{C}$ and contributions of deep water below 80 cm depth ($R^2 = 0.42, p < 0.001$) (Fig. 9b).
393 These linear relationships indicated that increasing reliance on upper soil water sources
394 linearly declined the leaf $\delta^{13}\text{C}$ and WUE of riparian trees (Fig. 9a), while the growing
395 dependence on deeper water sources proportionally increased the WUE of riparian trees (Fig.

396 9b). Riparian trees mainly relied on upper soil water sources fed by recent precipitation in
397 wet season (Dawson and Pate, 1996; Wang et al., 2019a). The easier soil water accessibility
398 stimulated riparian trees to maximize their growth and transpiration during the rainy season,
399 which resulted in the lower WUE and water use without water limitation (Cao et al., 2020;
400 Wu et al., 2021). Deep water was critical for plant water use and growth especially under
401 drought stress (Cao et al., 2020). However, the main root water uptake depth shift from
402 shallow layers towards deep layers (i.e., increasing reliance on deeper water sources)
403 generally caused riparian trees to consume more energy due to new deep-root growth, longer
404 water transport path and greater water transport resistance of deep-roots (Cao et al., 2020).
405 Therefore, riparian trees used deeper water effectively and conservatively through stomatal
406 closure and the lowering of transpiration rate, leading to higher leaf $\delta^{13}\text{C}$ and WUE values. It
407 agreed well with previous studies that the root water uptake depth adjustments towards
408 deeper layers during dry season could cause a 13% reduction of the total annual transpiration
409 and a 15% increase of the WUE of evergreen *Ligustrum lucidum* (Wu et al., 2021).

410 <Figure 9>

411 *4.3 Implications and further scopes of this study*

412 This study indicated that riparian deep-rooted trees were strongly dependent on shallow
413 soil water recharged by precipitation and remained high transpiration rates (low WUE) under
414 deep WTD. The continuously large plant water consumption could result in a dried upper soil
415 layer especially during drought season, increasing the risk of canopy defoliation, mortality
416 and degradation of riparian trees (Nie et al., 2014; Wang and D'Odorico, 2019; Ding et al.,
417 2020). Large-scale afforestation will not be recommended in this area, while shallow-rooted

418 herbs with less vulnerable to drought and lower water demand for transpiration may be
419 suitable for vegetation restoration in the riparian zone. A shift of shallow and deep-root water
420 uptake depths with higher WUE could help riparian trees acclimate to decreasing deep water
421 accessibilities under fluctuated shallow WTD (Dawson and Pate, 1996; Wang et al., 2019a).
422 Moreover, the higher WUE of riparian trees was generally linked to reducing water
423 consumption for transpiration without affecting plant growth and maintaining/reserving
424 riverflow and groundwater level, which benefited for balancing the growth of riparian trees and
425 bank storage to avoid loss of riverflow (Vanderklein et al., 2014; Barbeta et al., 2015; Missik et
426 al., 2019; Mkunyana et al., 2019). When the water table continuously declined and WTD
427 exceeded 2.1 m, deep water contributions to riparian trees as well as WUE decreased,
428 indicating that the ability to cope with drought stress decreased. In comparison, deep-root
429 development of trees was inhibited and photosynthetic/transpiration decreased due to
430 anaerobiosis when the water table kept high without evident fluctuation. Therefore, the
431 optimum WTD of 2.1 m for best growth of riparian trees was recommended under shallow
432 WTD condition. The sustainability of riparian afforestation must consider both root water
433 uptake depths and plant physiological traits.

434 This study provides a critical insight into evaluating the riparian plant-water relations
435 along a losing river. However, several issues still require further investigations. Firstly, the
436 stable isotopes should be combined with measurements of root distribution and transpiration
437 rate under more water table conditions in different represent years (i.e., dry, normal and wet
438 year) to further analyze the plant-water relationships. Secondly, whether there is a water use
439 competition between trees, shrubs and herbs in the riparian zone is unclear. It is valuable to

440 compare the water use characteristics among various riparian species in response to different
441 losing river water-groundwater interactions. Thirdly, numerical simulation model can be
442 combined with stable isotopes to trace the transport of water isotopes in
443 groundwater-soil-plant-atmosphere continuum and determine the dynamic characteristics of
444 plant water use (Simunek and Hopmans, 2009; Zhou et al., 2021).

445 **5. Conclusions**

446 In this study, the water use characteristics of deep-rooted *S. babylonica* in the riparian
447 zone under various WTDs along a losing river were determined by the stable isotopes ($\delta^2\text{H}$
448 and $\delta^{18}\text{O}$ in water as well as leaf $\delta^{13}\text{C}$) and MixSIAR model. Our results indicated that the
449 seasonal water uptake patterns and WUE of riparian *S. babylonica* were significantly affected
450 by WTD. Riparian trees accessed more upper soil water and had lower leaf $\delta^{13}\text{C}$ under deep
451 WTD (about 20 m) compared to those under shallow WTD (< 2.5 m). A shift of shallow and
452 deep root water uptake depths with higher WUE helped riparian trees adapt decreasing deep
453 water accessibilities under fluctuated shallow WTD (1.9 ± 0.4 m). Nevertheless, deep-root
454 water uptake capabilities and transpiration rate were limited when the water table kept high
455 without evident fluctuation (1.5 ± 0.1 m). Both the water source contributions in upper layer
456 (0–80 cm) and deep layer (below 80 cm depth) were parabolic functions of WTD ($p < 0.05$)
457 when groundwater was shallow. Leaf $\delta^{13}\text{C}$ had a positive linear correlation with proportional
458 contribution of deep water ($p < 0.001$). It seemed that 2.1 m was the optimum WTD for
459 riparian trees under shallow WTD, since they could make best use of deep-water sources and
460 remained the highest WUE under water deficit. The higher WUE of riparian trees was
461 advantageous for balancing plant water use and bank storage to avoid large loss of river flow.

462 However, large-scale afforestation will not be recommended in the area with deep WTD,
463 which may result in a dried upper soil layer especially during drought season and increase the
464 risk of degradation of riparian trees. The sustainability of riparian vegetation restoration
465 should consider both the development of root water uptake depths and physiological traits of
466 plants under different water table conditions.

467

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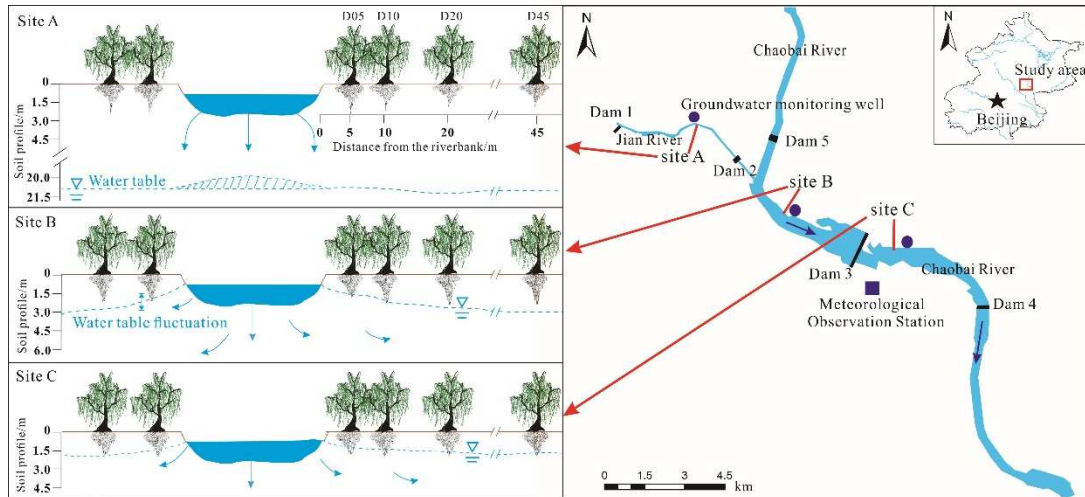


Fig. 1. Location of the study area and three representative sampling sites (A, B, and C). D05, D10, D20, and D45 at each site are the plots at distance of 5 m, 10 m, 20 m, and 45 m away from the riverbank, respectively.

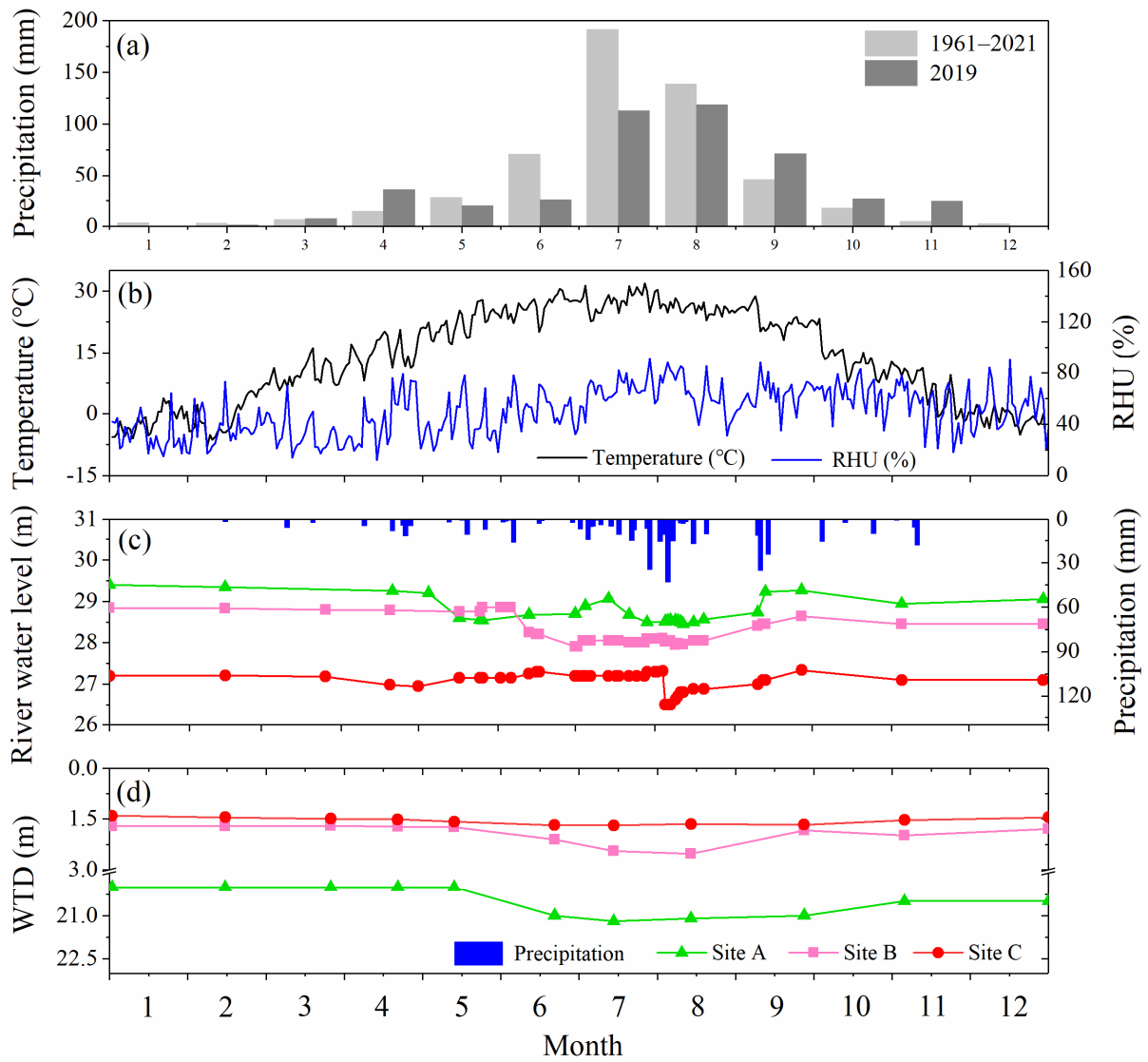


Fig. 2. (a) Monthly average precipitation for multi-year (1961–2021) and the observation period (2019), diurnal variation in (b) temperature and relative humidity (RHU), and seasonal changes in (c) river water level and precipitation, and (d) water table depth (WTD) at sites A, B, and C in 2019.

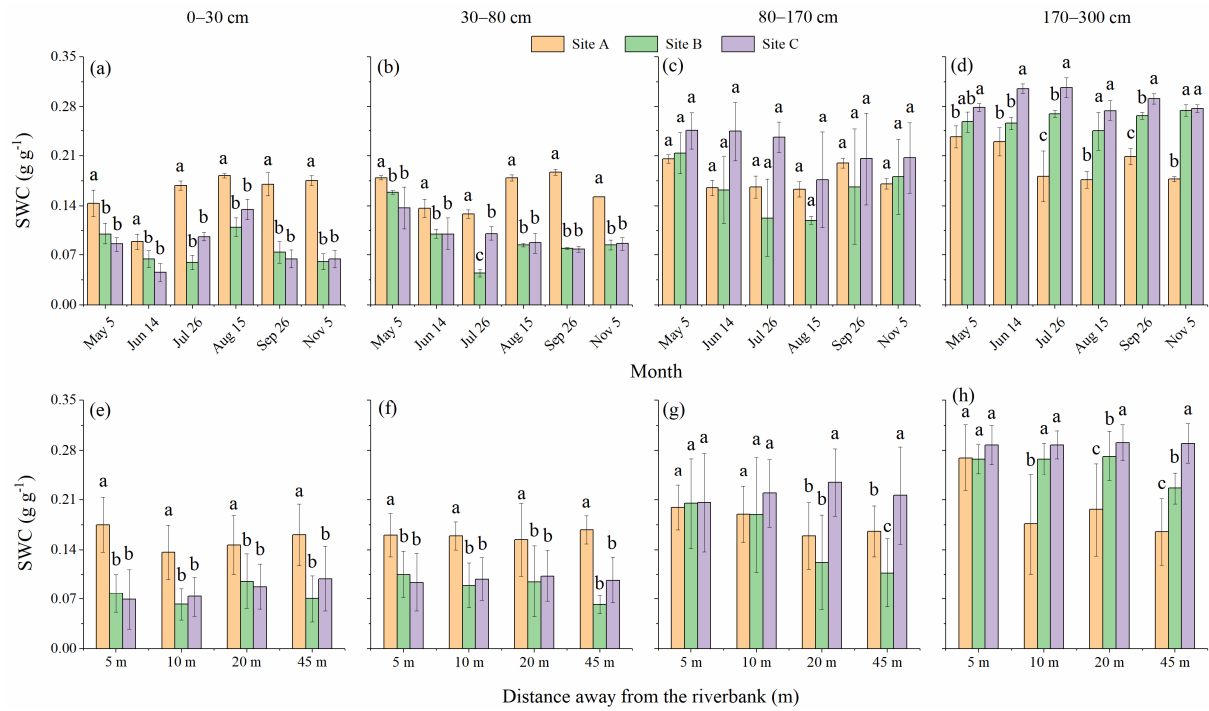


Fig. 3. Comparison of soil water content (SWC) in 0–30 cm, 30–80 cm, 80–170 cm and 170–300 cm layers (a–d) on six campaigns on average of four distances from the riverbank, and (e–h) in the four plots at distance of 5 m, 10 m, 20 m, and 45 m away from the riverbank on average of six campaigns during the observation period in 2019 among sites A, B, and C. The letters (a, b and c) represent significant differences of SWC among the three sites ($p < 0.05$).

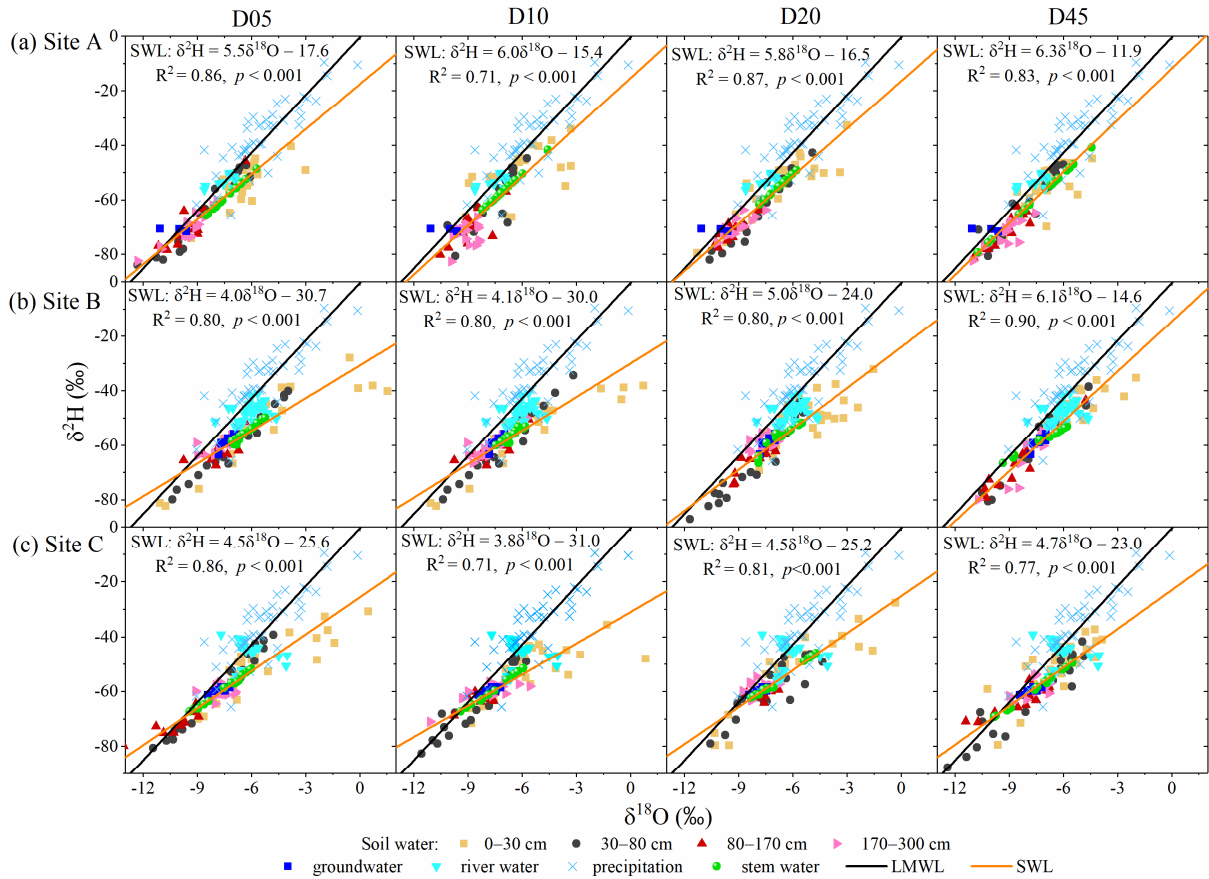


Fig. 4. Dual-isotope ($\delta^2\text{H}$ and $\delta^{18}\text{O}$) biplots of different water bodies in four plots (D05, D10, D20, and D45) during the observation period in 2019 at sites (a) A, (b) B, and (c) C. The soil water line (SWL) was fitted by $\delta^2\text{H}$ and $\delta^{18}\text{O}$ of soil water at different depths. The Local Meteoric Water Line (LMWL) fitted by the isotopic compositions of precipitation in 2019 in the study area was established as $\delta^2\text{H} = 5.5\delta^{18}\text{O} - 7.9$ ($R^2 = 0.81$, $p < 0.05$). Error bars are standard deviations. D05, D10, D20, and D45 are the plots at distance of 5 m, 10 m, 20 m, and 45 m away from the riverbank, respectively.

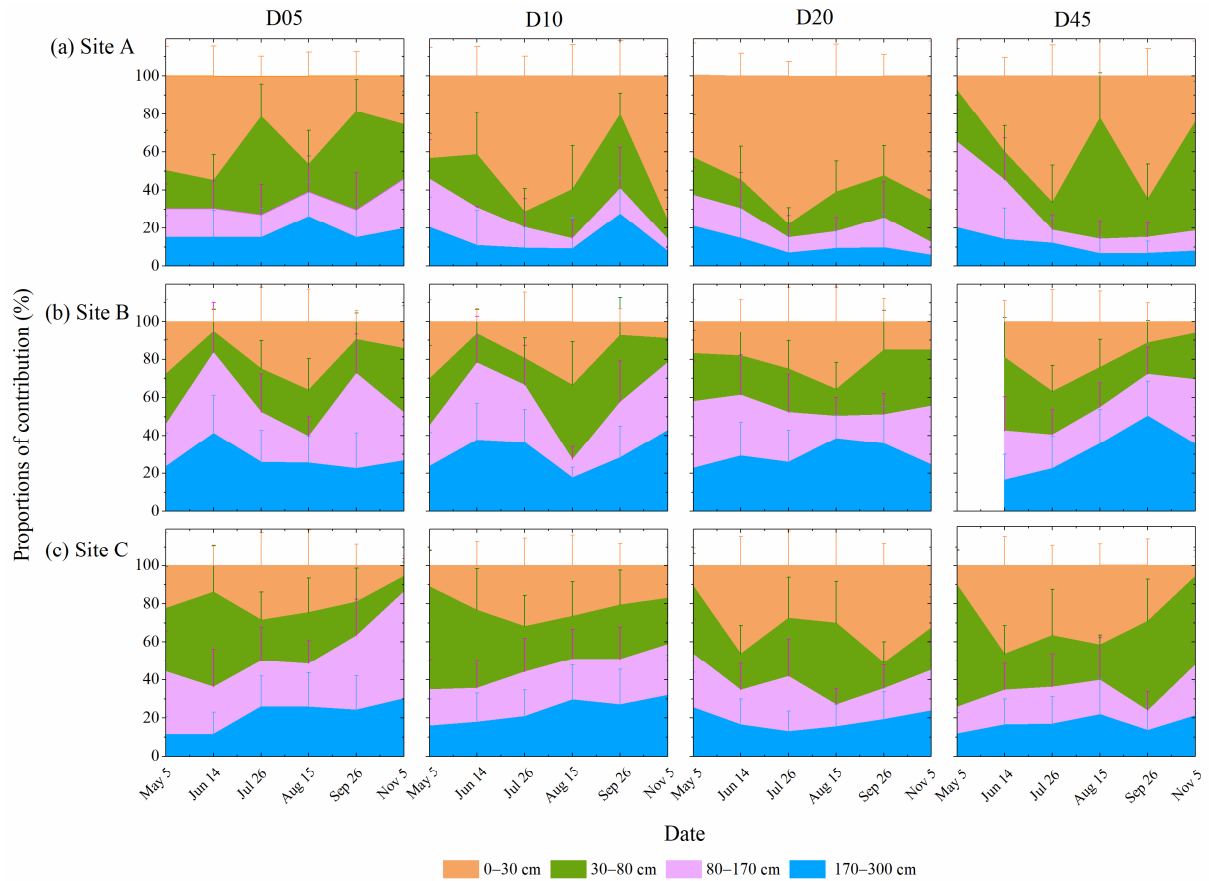


Fig. 5. Seasonal variations in the water uptake patterns of riparian *S. babylonica* in four plots (D05, D10, D20, and D45) at sites (a) A, (b) B, and (c) C. D05, D10, D20, and D45 are the plots at distance of 5 m, 10 m, 20 m, and 45 m away from the riverbank, respectively. The isotopic compositions of soil water in 170–300 cm layer also represent those of groundwater at sites B and C.

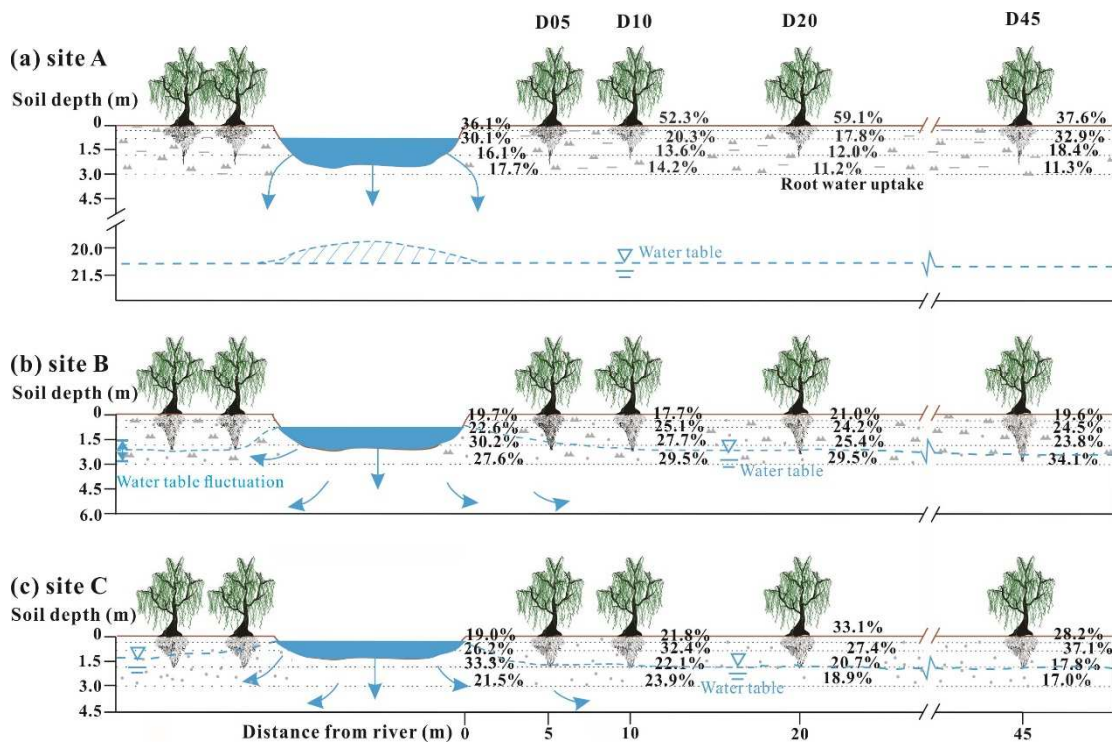


Fig. 6. Schematic diagram for the average root water uptake patterns of riparian *S. babylonica* for all six campaigns during the observation period in 2019 in four plots (D05, D10, D20, and D45) at sites (a) A, (b) B, and (c) C. D05, D10, D20, and D45 are the plots at distance of 5 m, 10 m, 20 m, and 45 m away from the riverbank, respectively.

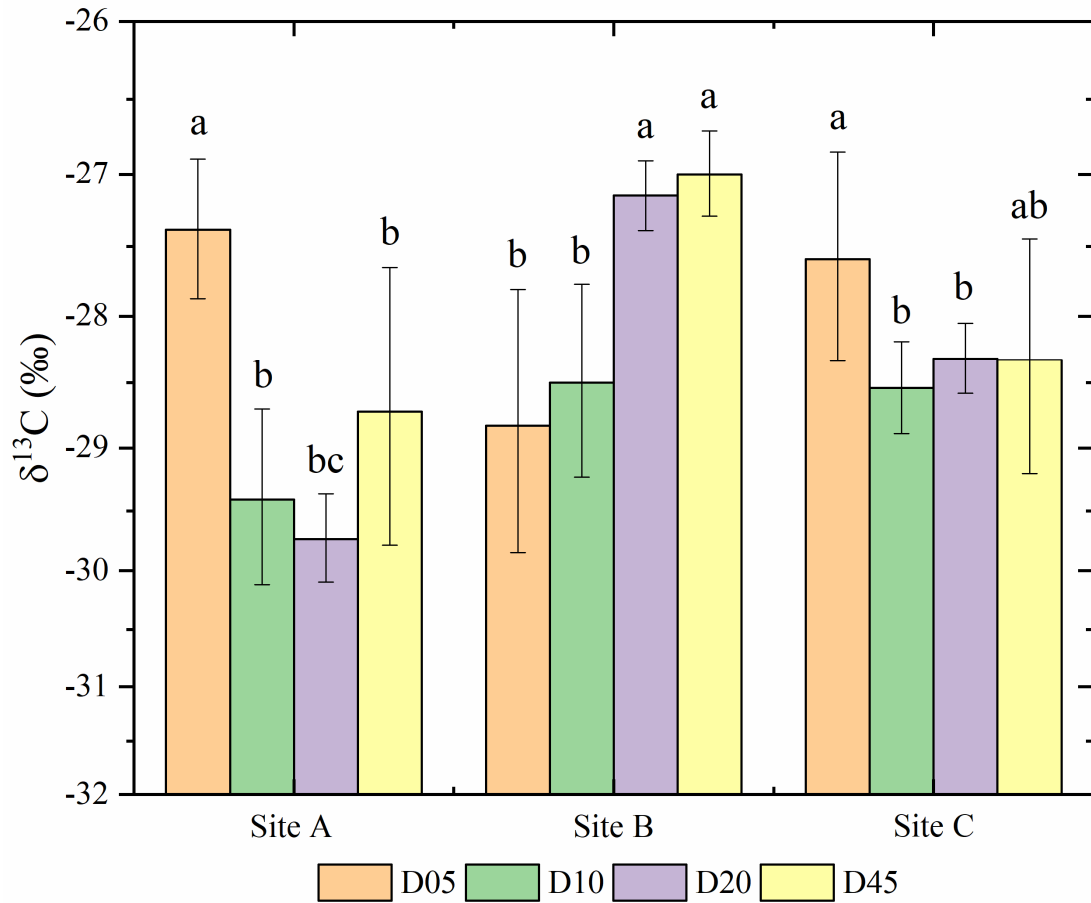


Fig. 7. The average leaf $\delta^{13}\text{C}$ of riparian *S. babylonica* for all six campaigns during the observation period in 2019 in four plots (D05, D10, D20, and D45) at sites A, B, and C. D05, D10, D20, and D45 are the plots at distance of 5 m, 10 m, 20 m, and 45 m away from the riverbank, respectively. The letters (a, b and c) represent significant differences in leaf $\delta^{13}\text{C}$ of riparian *S. babylonica* among the four plots (D05, D10, D20, and D45) ($p < 0.05$).

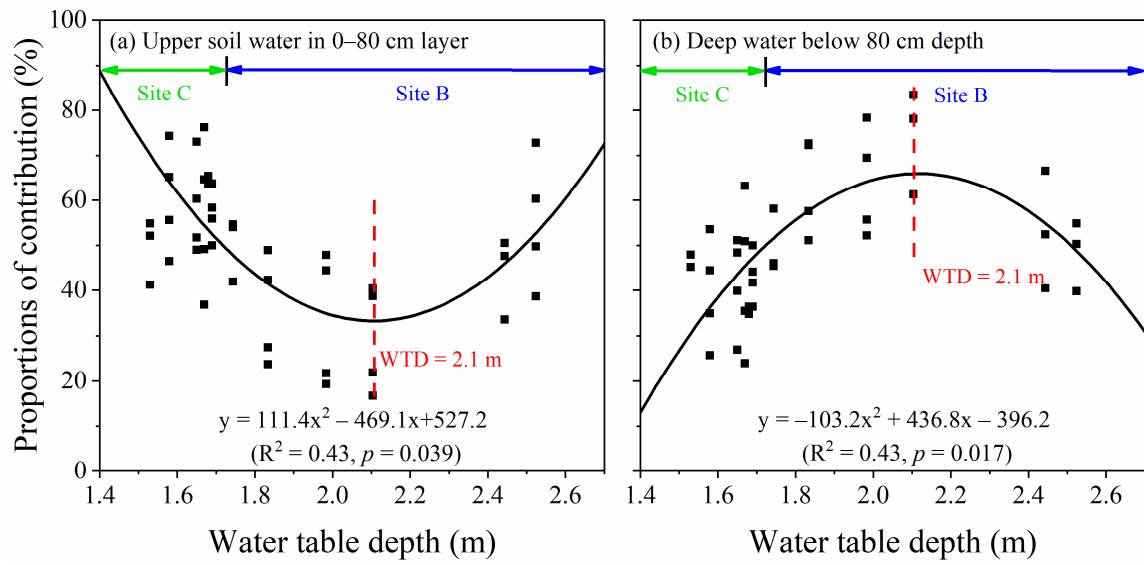


Fig. 8. Relationships between water table depth (WTD) and proportional contributions of (a) upper soil water in 0–80 cm layer and (b) deep water below 80 cm depth for riparian *S. babylonica* under shallow WTD conditions at sites B and C.

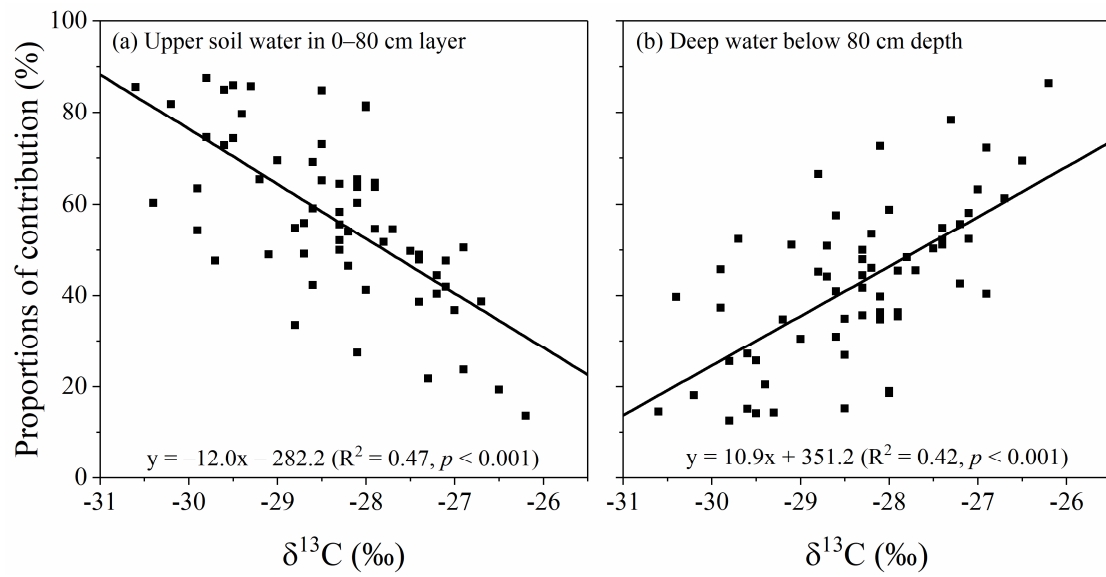


Fig. 9. Relationships between leaf $\delta^{13}\text{C}$ and proportional contributions of (a) upper soil water in 0–80 cm layer and (b) deep water below 80 cm depth for riparian *S. babylonica* at the three sites.